

1   **Title:**

2   **Ecological mechanisms explaining interactions within plant-hummingbird networks:**  
3   **Morphological matching increases towards lower latitudes**

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74

75    **Abstract**

76    Interactions between species are influenced by different ecological mechanisms, such as  
77    morphological matching, phenological overlap, and species abundances. How these mechanisms  
78    explain interaction frequencies across environmental gradients remains poorly understood.  
79    Consequently, we also know little about the mechanisms that drive the geographical patterns in  
80    network structure, such as complementary specialization and modularity. Here, we use data on  
81    morphologies, phenologies and abundances to explain interaction frequencies between  
82    hummingbirds and plants at a large geographic scale. For 24 quantitative networks sampled  
83    throughout the Americas, we found that the tendency of species to interact with morphologically  
84    matching partners contributed to specialized and modular network structures. Morphological  
85    matching best explained interaction frequencies in networks found closer to the equator and in areas  
86    with low temperature seasonality. When comparing the three ecological mechanisms within  
87    networks, we found that both morphological matching and phenological overlap generally  
88    outperformed abundances in the explanation of interaction frequencies. Together, these findings  
89    provide insights into the ecological mechanisms that underlie geographical patterns in resource  
90    specialization. Notably, our results highlight morphological constraints on interactions as a potential  
91    explanation for increasing resource specialization towards lower latitudes.

92  
93    **Keywords:** resource specialization, forbidden links, modularity, morphological matching,  
94    phenology, abundances, pollination.

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## 98    **Introduction**

99    The mechanisms that underlie the structure of interaction networks is a frequently investigated topic  
100    in community ecology [1-3]. For mutualistic networks, these mechanisms often comprise  
101    morphologies, phenologies, and abundances, which all may influence the likelihood for species to  
102    interact [1, 2, 4-8]. Knowledge about which mechanisms determine interaction frequencies has  
103    implications for interpreting the drivers of resource specialization [9-12]. However, to date, only  
104    studies of single networks have addressed the relevance of morphologies, phenologies, and  
105    abundances for explaining interaction frequencies between mutualistic partners [6, 13-17].  
106    Consequently, we still lack knowledge about how these ecological mechanisms may cause  
107    variability in resource specialization at large geographic scales.

108        Morphological matching is a frequently discussed determinant of mutualistic interactions [1,  
109    5, 14, 17, 18]. In plant-pollinator communities, morphological matching often represents the  
110    similarity in length and shape between a flower's corolla and a pollinator's feeding apparatus [4, 19,  
111    20]. Matching between these traits may lead to specialized network structures by limiting the  
112    species' access and preferences to a subset of mutualistic partners. Phenological overlap between  
113    interacting species is a mechanism that limits the available time for interactions to occur [2, 21].  
114    Seasonal turnover in community composition may result in some mutualists having none or limited  
115    co-occurrence, and consequently, few possibilities to interact [2, 21, 22]. By constraining  
116    interaction probabilities, phenological mismatch is expected to increase a network's level of  
117    specialization. Lastly, the local variation in species abundances may influence the random chance  
118    for partners to interact [2, 3, 23, 24]. Specifically, under random encounters, the likelihood of  
119    interactions would be highest between abundant species [1, 5]. Interaction frequencies that reflect  
120    abundances would thereby predict a high overlap between the resource use of species, resulting in a  
121    low degree of network specialization [16, 18, 23].

122       The degree to which morphological matching, phenological overlap, and abundance explain  
123 interaction frequencies may depend on the environmental setting. The classical literature on  
124 diversity gradients suggests that the warm, humid and less seasonal climates of the tropics make  
125 natural selection more influenced by biotic interactions in comparison to temperate regions [25-28].  
126 In turn, tropical climates could lead to a greater variety of morphological (co)adaptations and  
127 greater resource specialization [10, 25-28]. Accordingly, studies have found plant-hummingbird  
128 networks to become increasingly specialized towards the equator [10] and to coincide with greater  
129 diversity of hummingbird functional traits [8]. On this basis, one may hypothesize that the increase  
130 in resource specialization towards tropical regions coincides with an increased tendency of species  
131 to interact with morphological matching partners.

132       In contrast to morphological matching, to the best of our knowledge, there are no clear  
133 hypotheses on how phenological constraints on interactions vary across environmental gradients.  
134 Plant species in seasonal tropical regions usually have synchronized flowering periods during the  
135 rainy season, whereas phenologies are more staggered across the year in less seasonal regions [29].  
136 If the phenologies of pollinator species follow the same pattern, mutualistic partners in seasonal  
137 environments should generally experience few phenological mismatches. In that case, we would  
138 expect phenologies to impose fewer constraints on interaction frequencies in seasonal environments  
139 than in less seasonal environments. If both morphological and phenological constraints on  
140 interactions are absent, we expect that species would interact randomly according to their relative  
141 abundances [18]. Thus, abundance should best explain interaction frequencies in cold, dry and  
142 seasonal environments [3, 24].

143       In this study, we use morphological matching, phenological overlap, and abundance to  
144 explain plant-hummingbird interaction frequencies at a large geographic scale. We focus our  
145 investigation on these three mechanisms, as they have previously been shown to influence

146 interaction frequencies within plant-pollinator networks [2, 5, 6, 14, 16]. First, we investigate how  
147 the mechanisms may influence network-derived measures of resource specialization, and second,  
148 how their explanation of interaction frequencies varies across environmental gradients (see ESM1  
149 for a conceptual overview of our main hypotheses). To do so, we analyzed a unique dataset  
150 consisting of 24 plant-hummingbird interaction networks distributed from Central Mexico to  
151 Southern Brazil (20°N-30°S). For each plant and hummingbird species within these networks, we  
152 have additional information on phenologies, morphologies, and abundances. We expected  
153 morphological matching and phenological overlap to be more relevant for explaining interaction  
154 frequencies in regions with low annual climate seasonality and high annual mean temperature and  
155 precipitation [8, 10-12, 25, 26, 30]. It has been proposed that specialized co-adaptations increase in  
156 tropical mountains due to the suitable preconditions for species to track changing climates [11, 31,  
157 32]. Thus, we expected that morphological matching and phenological overlap would impose more  
158 constraints on interaction frequencies in topographically complex areas in the tropics. Oppositely,  
159 we expected abundance to have a decreased influence on interaction frequencies in specialized  
160 networks sampled in areas with high topographical complexity, high mean annual temperature, high  
161 mean annual precipitation, and low climate seasonality [16, 18].

162

## 163 **Methods**

### 164 *(a) Networks, abundance, phenology and species traits*

165 We compiled data on 24 quantitative interaction networks collected throughout the Americas, in  
166 areas mostly or entirely covered with native vegetation [Table ESM2; updated dataset from 24].  
167 The networks comprise 106 hummingbird species, 31% of all described hummingbird species in the

168 world according to the IOC World Bird List v.7.3 [33; ESM3], and 450 plant species belonging to  
169 57 plant families (ESM4, see ESM5a for additional details on sampling).

170 The abundance of plant species was measured as the number of flowers produced per species  
171 in each community throughout the entire sampling period. Flowers were counted in plots or  
172 transects estimated regularly throughout the sampling period. The abundance of hummingbirds  
173 within sites was measured in the field by counting the number of visual and aural detections of  
174 individuals across transects (n=12 networks) or point counts (n=4 networks), or the number of  
175 individuals captured by mist-netting (n=8 networks; ESM5a). Because the abundance sampling  
176 protocols were not standardized among networks, we treated the data as relative abundance, i.e. for  
177 all species we calculate their abundance as the proportion of the total number of individuals within a  
178 given community. Still, we note that mist nets may be especially efficient for surveying elusive  
179 understory species, such as traplining hummingbirds, whereas transects and point counts may be  
180 better at surveying species at higher vegetation strata [34]. We recognize that the caveat inherent in  
181 using different sampling schemes across networks may influence the outcome of our analyses.  
182 However, as we used relative abundances to model interaction frequencies within networks (not  
183 between networks), we believe that the different sampling schemes had a minimal influence on our  
184 results.

185 The phenology of each plant and hummingbird species in each network was determined as the  
186 presence-absence of, respectively, open flowers and individuals at each sampling period (usually  
187 months). Flower morphology was characterized by the effective corolla length [sensu 35], measured  
188 as the distance from the nectary to the corolla opening. The effective corolla length reflects the  
189 minimum length of mouthparts required for pollinators to access the nectar legitimately.  
190 Hummingbird bill morphology was measured as the length of the exposed culmen from captured  
191 hummingbird individuals (see ESM5b further details on sampling).



192

193 *(b) Climate and topography data*

194 All climate variables were extracted as means within a 10 km radius around the location of each  
195 network. Climate variables were downloaded from the WorldClim database in 30 arc second  
196 resolution [36; v 2.0; <http://www.worldclim.org>]. We included mean annual temperature (MAT)  
197 and precipitation (MAP), as well as seasonality in temperature (annual standard deviation in  
198 monthly mean temperature; TS) and precipitation (annual coefficient of variation in monthly  
199 precipitation; PS). Topographical data were retrieved from SRTM 90m DEM Version 4  
200 (<http://srtm.csi.cgiar.org>). Within a 10 km zone around each network's location, we determined  
201 topographical range as the difference between the maximum and minimum elevation.

202 *(c) Defining models for morphological matching, phenological overlap, and abundance*

203 We used the model for morphological matching presented in Sonne *et al.* [37], which assumes  
204 pairwise interaction frequencies to decrease proportionally to the standardized difference in length  
205 between the hummingbird's bill and the plant's floral corolla. In this framework, a hummingbird  
206 with the longest bill has the highest probability of interacting with the longest flowers, while the  
207 shortest-billed hummingbirds have the highest probability of interacting with the shortest flowers.  
208 We calculated pairwise morphological matching for each hummingbird  $i$  and plant  $j$  within each  
209 network  $k$ . We first standardized the hummingbird's bill length and the plant's corolla length to zero  
210 mean and unit variance within networks and then calculated the Euclidean distance between each  
211 pair of species ( $M_{ijk}$ ; see ESM6a for details).

212 In the phenological overlap model, pairwise interaction frequencies increase proportionally  
213 to the number of sampling periods (usually months) in which hummingbird  $i$  and flowering plant  
214 species  $j$  co-occurred in network  $k$  [ $P_{ijk}$ ; 2]. Random species encounters should generate pairwise

215 interaction frequencies that are proportional to the partners' relative abundances [2]. Therefore, the  
216 pairwise interaction probabilities in network  $k$  are calculated by multiplying the relative abundances  
217 of hummingbird  $i$  and plant  $j$  ( $A_{ijk}$ ). Prior to statistical modelling,  $M_{ijk}$ ,  $P_{ijk}$ , and  $A_{ijk}$  were scaled to  
218 zero mean and unit variance.

219  
220 *(d) Linking ecological mechanisms with species' interaction frequencies*

221 We used morphological matching ( $M$ ), phenological overlap ( $P$ ) and species abundances ( $A$ ) to  
222 model interaction frequencies individually for each hummingbird and plant species. The modeling  
223 procedure was modified from Weinstein & Graham [38] as our data do not allow the estimation of  
224 interaction detectability. While the following model is described for hummingbirds, the same  
225 method was applied also to the plants.

226 The model assumes that the pairwise interaction frequency ( $Z$ ) for each hummingbird  
227 species  $i$  in the network  $k$  follow a Poisson distribution with mean  $\lambda_{ik}$ , with the log link function of  
228  $\lambda_{ik}$  predicted by the covariates:  $M_{ijk}$ ,  $P_{ijk}$ , and  $A_{ijk}$ . Model parameters were estimated using Markov  
229 chain Monte Carlo (MCMC). From there, we extracted posterior distributions of the standardized  
230 coefficients:  $\beta^M_{ik}$ ,  $\beta^A_{ik}$ , and  $\beta^P_{ik}$ . As values of  $M$  are distances measuring morphological mismatch, a  
231 more negative  $\beta^M_{ik}$  value indicates greater ability of morphological matching to explain interaction  
232 frequencies. Conversely, more positive  $\beta^A_{ik}$  and  $\beta^P_{ik}$  values indicate better explanations of  
233 interaction frequencies by abundances and phenological overlap. These parameters were considered  
234 significant if less than 5% of their estimated posterior distribution overlapped with zero [38; see  
235 ESM6b for details on the model specification]. For each network, we determined the proportion of  
236 species with interaction frequencies significantly explained by morphological matching ( $Imp.M$ ),  
237 phenological overlap ( $Imp.P$ ) and abundance ( $Imp.A$ ). We used these proportions as measures of  
238 each mechanisms' ability to explain the hummingbirds' interaction frequencies. We also calculated

the *relative* proportion of species by dividing *Imp.M*, *Imp.P*, and *Imp.A* by their within-network sum. Thus, the relative proportions value represents the degree to which a given ecological mechanism performs in explaining interaction frequencies relative to the remaining two mechanisms. We present relative proportion values for each network as kernel density distributions [39]. Here, the contribution of each network is weighted according to the richness of hummingbirds.

#### (e) *Linking ecological mechanisms with network structure*

We investigated whether morphological matching, phenological overlap, and abundances influenced two measures of network structure that reflect resource specialization. First, we calculated *complementary specialization* ( $H_2'$ ), which quantifies the partitioning of interactions relative to their availability [i.e. network marginal sums; 40]; it ranges between zero and one, with higher values indicating higher partitioning of interactions between species in the community. Second, we calculated *weighted modularity* ( $Q$ ), which describes a network's tendency to comprise subgroups of interacting species [9]. Modularity is high when species mostly interact with partners from their respective modules. Modularity differs from specialization by quantifying the partitioning of interactions between groups of species rather than between individual species. Modularity was calculated using the DIRTLP+ algorithm [41] using  $10e^7$  steps. Due to the stochastic nature of this optimization algorithm, we repeated the analysis 10 times for each network and kept the highest  $Q$  value [9].

As empirical networks vary in species richness and sampling intensity, the observed values of network structure, such as specialization and modularity may not be directly comparable [42]. To overcome this issue, we used null models to calculate the extent to which the observed network metric deviates from a null expectation. To compute the null model, we used Patefield's algorithm [43], and then subtracted the observed values of network structure from the mean values obtained from 1000 randomizations ( $\Delta$ -transformation; See ESM7 for evaluation of alternative null models).

263 While we report the null model corrected values for modularity [40, 44], we report complementary  
264 specialization without the null model correction because this index is already subject to a correction  
265 for the marginal totals of the network. Nevertheless, we also ran analyses with the  $\Delta$ -transformed  
266 version as reported in the supplementary material. All network analyses were conducted in R using  
267 the ‘bipartite’ package v. 2.11 [45].

268 We fitted linear models for complementary specialization and  $\Delta$ modularity using the  
269 proportion of species within networks whose interaction frequencies were significantly explained  
270 by morphological matching (*Imp.M*), phenological overlap (*Imp.P*) and abundance (*Imp.A*). As  
271 additional explanatory variables, we included the network size, defined as the total richness of  
272 hummingbird and plant species, and a measure of sampling intensity, calculated as the square root  
273 number of interaction events divided by the total richness of hummingbird and plant species [46,  
274 47]. We also tested for the potential confounding effect of spatial autocorrelation in the linear model  
275 residuals using the R package *ncf* [48]. To do this, we fitted Moran’s *I* correlograms with 500 km  
276 distance classes and a truncation distance of 5000 km. Positive spatial autocorrelation was non-  
277 significant in all models, suggesting that spatial autocorrelation had no influence on our results  
278 (ESM8).

279

#### 280 *(f) Geographical patterns in ecological mechanisms’ explanation of interaction frequencies*

281 Separately for hummingbirds and plants, we fitted logistic models that regressed absolute latitude  
282 against *Imp.M*, *Imp.P*, *Imp.A*. To explore the effect of species richness and sampling, we ran  
283 supplementary models that included network size (total richness of hummingbird and plant species)  
284 and sampling intensity as explanatory variables (ESM9). We noticed that networks were not evenly  
285 sampled across the Americas, with the majority of networks occurring either in Central America/the  
286 Andes or along the Brazilian Atlantic coast (ESM9). To assess if the latitudinal trend was caused by

287 differences between these two biogeographical regions, we regressed absolute latitude against  
288 *Imp.M*, *Imp.P*, and *Imp.A* while including a dummy variable stating whether networks were located  
289 east or west of 60° longitude (i.e., occurred in Central America/Andes or along the Brazilian  
290 Atlantic coast).

291 Finally, we used logistic models to regress *Imp.M*, *Imp.P*, and *Imp.A* against network size and  
292 the following environmental variables: topographic range, mean annual temperature, mean annual  
293 precipitation, temperature seasonality, and precipitation seasonality. All possible model  
294 combinations were fitted and then evaluated using the Akaike Information Criterion with correction  
295 for small sample sizes (AIC<sub>C</sub>). Usually, we found no single best model for the response variables, as  
296  $\Delta\text{AIC}_C$  was  $\leq 2.0$  between the model with the lowest AIC<sub>C</sub> and several other models [49].  
297 Therefore, we averaged the estimates from all possible model combinations weighted by Akaike  
298 weights ( $w_i$ ). Additionally, we present the summed Akaike weights for all models containing each  
299 explanatory variable,  $\Sigma w_i$  [49]. Model selection and model averaging were conducted using the  
300 ‘MuMIn’ package in R [50]. The goodness of fit for linear models was evaluated by McFadden’s  $R^2$   
301 [51].

302

## 303 **Results**

304 The proportion of species interacting with morphologically matching partners (*Imp.M*) correlated  
305 positively with both metrics of resource specialization (*complementary specialization* and  
306 *modularity*; Figure 1). These correlations remained significant when accounting for the potential  
307 confounding influence of sampling intensity and network size (ESM10).

308 The tendency of species to interact with morphologically matching partners increased towards  
309 tropical latitudes (Figure 2). This trend remained after accounting for the clustering in network  
310 sampling between the Central America/Andes and Eastern Brazil, and when including network size

311 and sampling intensity as additional explanatory variables (ESM 9). For hummingbirds, we found  
312 phenological overlap to have a stronger influence on interaction frequencies towards tropical  
313 latitudes, while species' abundances better explained interaction frequencies towards higher  
314 latitudes, when including region (i.e. Central America/Andes vs. Eastern Brazil), network size and  
315 sampling intensity as explanatory variables (ESM9). For plants, however, phenological overlap  
316 better explained interaction frequencies within networks from Central America/Andes than within  
317 networks from Eastern Brazil (ESM9).

318 For both plants and hummingbirds, the explanation of interaction frequencies by  
319 morphological matching correlated negatively with temperature seasonality, which was the only  
320 predictor variable that remained present in all best-fitting models ( $\Delta AICc < 2.0$ ; Table 1).  
321 Phenological overlap best explained the hummingbirds' interaction frequencies in areas with low  
322 temperature seasonality (Table 1). For plants, phenological overlap best explained interaction  
323 frequencies in areas with high topographic range, mean annual temperature, temperature  
324 seasonality, and low species richness (Table 1). For hummingbirds, abundance best explained  
325 interaction frequencies in areas with high seasonality in precipitation and low seasonality in  
326 temperature (Table 1). Abundance was best explaining the plants' interaction frequencies in areas  
327 with high topographic range, mean annual temperature and temperature seasonality (Table 1).

328 Overall, we found that abundance generally had a weaker influence on interaction frequencies  
329 than morphological matching and phenological overlap (Figure 3).

330

## 331 Discussion

332 Across plant-hummingbird networks, we show that resource specialization increases with the  
333 tendency of species to interact with morphologically matching partners (Figure 1). Moreover, we  
334 found that morphological matching is a more prominent driver of species' interaction frequencies

335 when the networks are located closer to the equator (Figure 2). Together, these results highlight that  
336 the relevance of traits for explaining interaction frequencies is context dependent [7]. In some  
337 communities, traits matter more than in others, and this variability may contribute to the latitudinal  
338 pattern in resource specialization. In addition to morphologies, we found phenologies to be  
339 important for determining interaction frequencies, more so than the local variability in abundance  
340 (Figure 3). Based on our findings we speculate that partitioning of ecological niches, by  
341 morphological matching and phenological overlap, could be important for maintaining coexistence  
342 within diverse communities of interacting species.

343       The fact that morphological matching best explained interaction frequencies closer to the  
344 equator is consistent with the classical view of tropical environments as favorable to biotic  
345 specialization, either through co-evolutionary adaptations or ecological fitting [10, 25, 52, 53]. Still,  
346 the idea that biotic specialization increases towards the tropics has remained a subject of debate, as  
347 empirical studies have presented idiosyncratic results [10, 46, 47, 54, 55]. Previous work on our  
348 hummingbird-plant networks showed that network-derived resource specialization increases  
349 towards tropical latitudes [10], and coincides with high diversity of hummingbird traits [8]. Our  
350 study goes beyond previous macroecological analyses of network structure [8, 10, 11], and targets  
351 the potential ecological mechanisms that may constrain the likelihood for partners to interact. We  
352 recognize the caveat inherent to the geographical distribution of our sampling localities (Fig. 2), and  
353 that expanding the sampling to cover a wider latitudinal gradient would be important to strengthen  
354 our conclusions. Yet, the results presented here bring us closer to a mechanistic understanding of  
355 how morphological traits may influence the variability in resource specialization across  
356 environmental gradients.

357       Beyond the latitudinal patterns, we also found that morphological matching best explained  
358 interaction frequencies in areas with low temperature seasonality (Table 1). Annual temperature

359 stability is one of the most important characteristics of tropical climates [56]. The lower seasonality  
360 of tropical regions causes resources to be present throughout the year, which may lay a foundation  
361 for plants and pollinators to develop specialized associations [11, 25, 26, 57]. Moreover, theory  
362 suggests competition to be more influential on community structure in the absence of environmental  
363 filters, as within the tropics [27, 30, 57]. Thus, we speculate that interspecific competition within  
364 guilds together with diffuse mutualistic coadaptations may contribute to the high degree of resource  
365 specialization within the tropical regions [8, 15, 58, 59]. Similar to morphological matching,  
366 phenological overlap had a stronger influence on hummingbird's interactions in areas with low  
367 temperature seasonality (Table 1). This result concurs with our hypothesis that seasonal climates  
368 cause more synchronized flowering periods among plant species in comparison to aseasonal  
369 climates. When plants synchronize their flowering, each pollinator species may overall experience  
370 fewer phenological mismatches with their mutualistic partners. For the plants, however, the same  
371 association pointed in the opposite direction: phenological overlap had a stronger influence on  
372 interactions in areas with a high temperature seasonality (Table 1). This result may be driven by  
373 hummingbirds with migratory behavior, which are more common in seasonal regions.  
374 Hummingbird migrations could impose higher seasonal turnover in species composition, thereby  
375 causing higher phenological constraints to the plants' interactions. Regardless of the underlying  
376 explanation, the idiosyncratic results between hummingbirds and plants highlight that the  
377 environmental conditions that cause seasonal species turnover may depend on the ecological guild.  
378 Contrarily to morphological matching and phenological overlap, our results do not support the idea  
379 that abundance is a general explanation for why some interactions occur more frequently than  
380 others. Thus, although abundant hummingbirds may be more generalized in their floral preferences  
381 [24], our result suggests that abundant species do not necessarily have the most frequent  
382 interactions with abundant partners. Thus, although phenology and to a lesser extent abundance also



383 matter, we have identified morphological matching as a key determinant of species' interaction  
384 frequencies.

385       The most challenging aspect of documenting morphological matching in plant-hummingbird  
386 networks is the fact that hummingbirds can extend their tongue to access the flowers' nectar. As  
387 such, the absolute difference between bill length and flower depth is not an accurate representation  
388 of the actual morphological mismatch. Our model for morphological matching builds on the  
389 assumption that hummingbirds with the relatively longest bills should prefer flowers with the  
390 longest corollas. Thereby the model anticipates morphological matching to affect the  
391 hummingbirds' floral preferences without making specific assumptions about how much their  
392 tongues can be extended. The model also implies that long-billed hummingbirds should interact less  
393 with short flowers [13]. The ecological explanation for this may be twofold. First, long-billed  
394 hummingbirds could minimize competition with short-billed species by using preferentially the  
395 flowers with the longest corollas [20]. Second, flowers with short corollas typically have low nectar  
396 volume, which could make them unprofitable for long-billed hummingbird [20, 60]. Therefore, both  
397 interspecific competition between hummingbirds and variability in the flowers' nectar reward may  
398 also explain why morphological matching contributes to a high degree of resource specialization  
399 [35, 60].

400       When compared to other avian pollination systems, hummingbirds and plants have the most  
401 specialized morphologies and the highest level of resource specialization [20, 61, 62]. Moreover,  
402 pollination networks are usually more specialized than seed dispersal networks, which is another  
403 common form of plant-bird mutualism [63]. In plant–frugivorous bird communities, network-  
404 derived specialization has previously been shown to increase towards temperate latitudes [47], but  
405 dietary specialization, i.e. the proportion of obligate frugivores, increases towards tropical latitudes  
406 [46]. As such, one may obtain opposing results depending on the scale at which resource

407 specialization is measured. While only some fruit-eating birds are specialized frugivores, all  
408 hummingbirds rely mostly on nectar as a food resource. Hence, plant-hummingbird networks  
409 should be a suitable system to investigate morphological matching. Although, many orders of insect  
410 pollinators have specialized adaptations to nectar extraction, they are also found collecting other  
411 floral resources [64]. In bees, for instance, tongue-lengths may affect floral preferences during  
412 nectar-gathering, but flowers with inaccessible nectar may still receive visits when bees are  
413 collecting pollen [64]. Plants and hawkmoths exhibit some highly specialized coadaptations, and  
414 this is probably the closest analog to a system in which morphological matching is crucial for  
415 determining interaction frequencies [65].

416       The mechanisms underlying the structure and organization of ecological communities have  
417 remained poorly understood at the large geographical scale. Using plant-hummingbird networks  
418 distributed across the Americas, we have identified morphological matching as a potential key  
419 driver of geographical patterns in resource specialization. Notably, closer to the equator, species  
420 tended to show stronger preferences for mutualistic partners with morphologically matching traits.  
421 Thus, our findings support the hypothesis that trait-mediated resource specialization increases  
422 towards lower latitudes.

423

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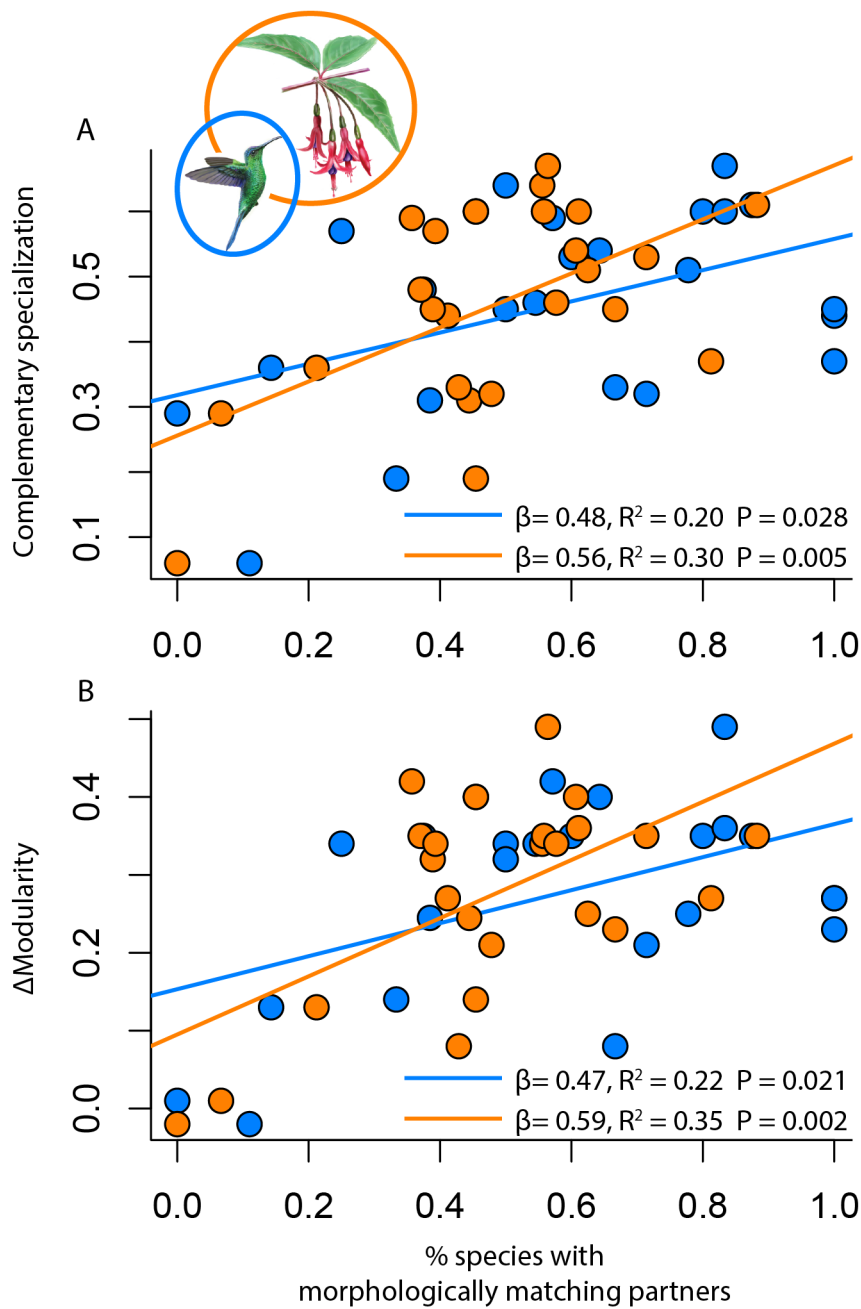
435   **Tables**

436   **Table 1.** Model selection and averaging results of logistic models fitted on different ecological  
437 mechanisms' ability to explain interaction frequencies within networks (n =24). The analyses were  
438 repeated individually for hummingbirds and plants. The response variables comprised the  
439 proportion of species within networks whose interaction frequencies were significantly explained  
440 by morphological matching (*Imp.M*), phenological overlap (*Imp.P*), and abundance (*Imp.A*).  
441 Explanatory variables include *Network size*: total richness of hummingbirds and plants within  
442 networks; *Topographic range*: maximum elevation subtracted from minimum elevation; *Mean*  
443 *annual temperature*; *Mean annual precipitation*; *Temperature seasonality*; and *Precipitation*  
444 *seasonality*. Goodness of fit is assessed by McFadden's  $R^2$ .  $\Sigma w_i$ : Sum of 'Akaike weights' from all  
445 models including the predictor variable. MAM: standardized coefficients of variables present in all  
446 minimum adequate models ( $\Delta AICc < 2$ ).  $N_{MAM}$ : number of minimum adequate models. AVM:  
447 standardized coefficients of the averaged model across all models including a focal predictor  
448 variable.

449

## 450   **Figures**

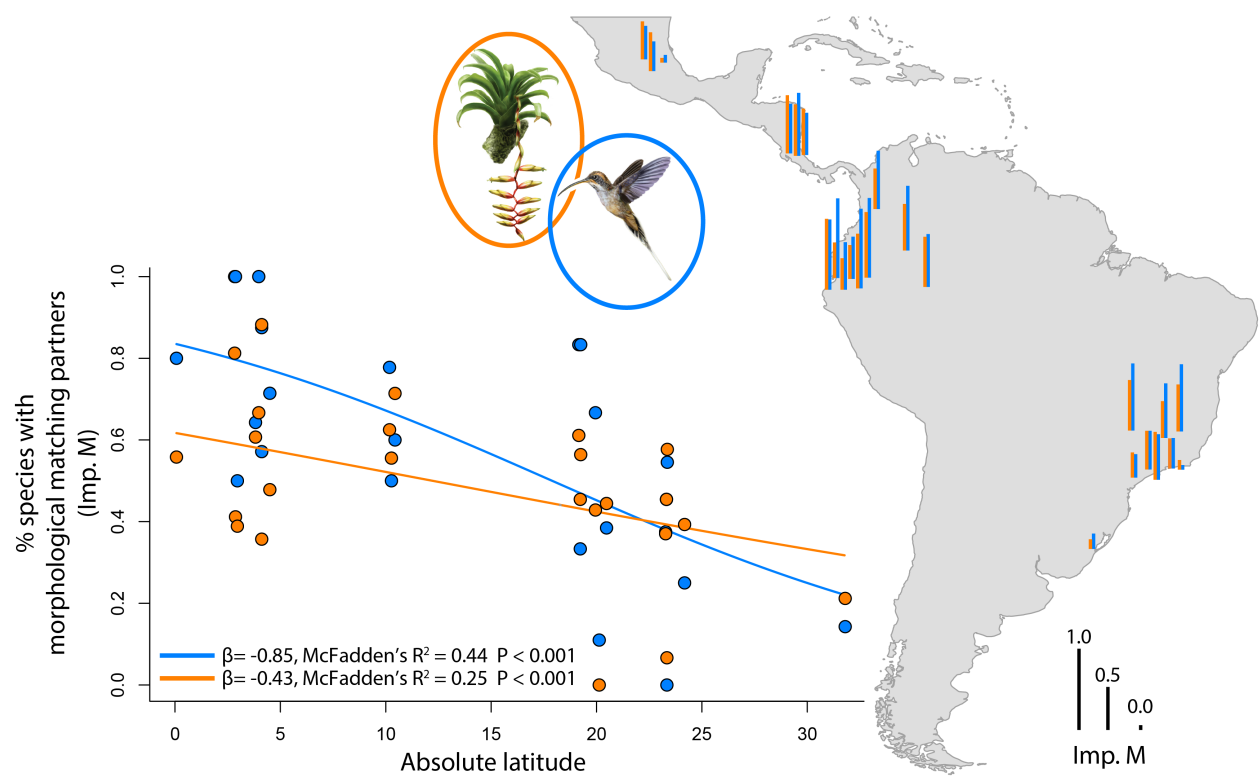
451   **Figure 1.** Linear models depicting the relationship between resource specialization and species'  
452   morphological matching. Resource specialization was measured as complementary specialization  
453   (A), and  $\Delta$ modularity (B).  $\Delta$  indicates a correction by the *Patefield* null model (see Methods). The  
454   x-axis shows the proportion of species within networks that exhibited a significant tendency to  
455   interact with morphologically matching partners. The linear models were run separately for  
456   hummingbirds (blue) and plants (orange). Drawings by Pedro Lorenzo.



457

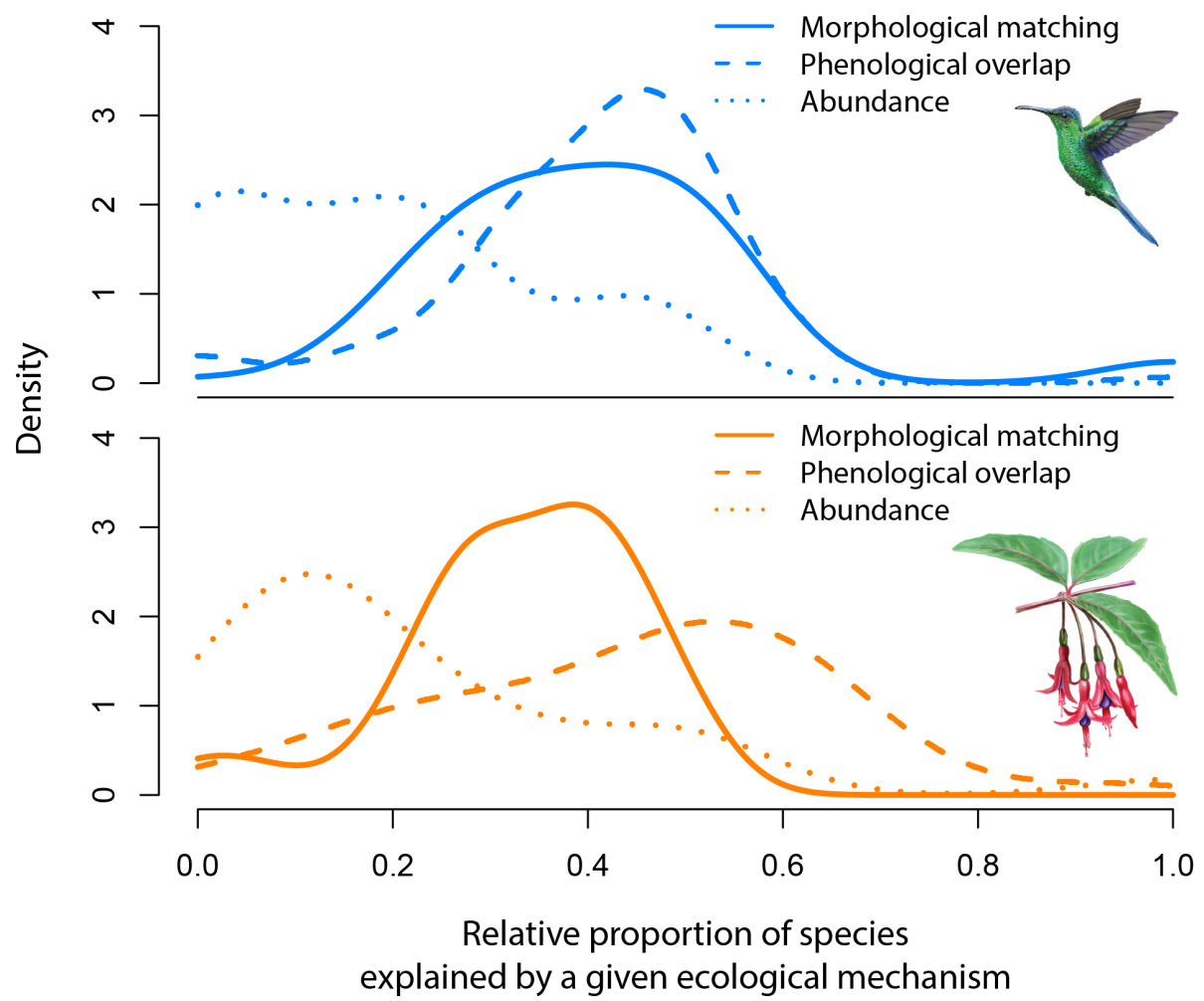
458 **Figure 2.** Geographical distribution of the 24 plant-hummingbird networks. The height of the bars  
 459 reflects the proportion of hummingbird species (blue) and plant species (orange) within networks  
 460 that showed significant tendencies to interact with morphologically matching partners (*Imp. M*).  
 461 Some points on the map have been slightly separated to improve visual clarity. The scatterplot

462 shows the negative relationship between *Imp. M* and absolute latitude, modeled by logistic  
 463 regression. Drawings by Pedro Lorenzo.



464

465 **Figure 3.** Kernel density distributions depicting the relative contribution of morphological  
 466 matching, phenological overlap, and abundance to the explanation of species' interaction  
 467 frequencies. The x-axis shows the proportion of hummingbirds (blue) and plants (orange) within  
 468 networks whose interaction frequencies are significantly explained by morphological matching,  
 469 phenological overlap and abundance, respectively. Each of these values was divided by their  
 470 within-network sum, thereby obtaining a *relative* proportion of species. Relative proportion values  
 471 higher than 0.50 indicate that a particular ecological mechanism was more effective at explaining  
 472 interaction frequencies than the two remaining mechanisms combined within a given network.  
 473 Drawings by Pedro Lorenzo.



474

475



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